

1 **Plant mercury pump controls seasonal variations in global atmospheric mercury**

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20 **Abstract**

21 Anthropogenic mercury emissions are transported through the atmosphere as gaseous
22 elemental mercury (Hg(0)) prior to deposition to earth surfaces. Two explanations for strong
23 Hg(0) seasonality in the Northern Hemisphere (NH) are variations in primary anthropogenic
24 Hg(0) emissions, thought to peak in winter due to higher energy consumption, and faster
25 atmospheric Hg(0) oxidation rates in summer. Global Hg models simulate equally
26 pronounced, oxidation-driven Hg(0) seasonality in the Southern Hemisphere (SH) where
27 Hg(0) levels are observed to be constant year-round. Furthermore, power consumption
28 related Hg emissions in N-America are higher in summer than in winter, which is inconsistent
29 with low N-American summertime Hg(0) concentrations. Here, we assess the role of Hg(0)
30 uptake by plants as an alternative driving mechanism for global Hg(0) variability. We find that
31 data from all terrestrial background sites in the NH show a co-variation of Hg(0) with CO₂,
32 which is known to exhibit concentration minima in summer due to CO₂ assimilation by
33 vegetation. The amplitude of seasonal Hg(0) oscillations increase with latitude in the NH and
34 are larger at inland terrestrial background sites compared to coastal sites. We find that
35 satellite-derived vegetation photosynthetic activity (NDVI) correlates with Hg(0) levels at
36 individual sites and across continents. These findings suggest vegetation to be a global Hg(0)
37 pump, that dominates seasonal variations of atmospheric Hg(0). We suggest that decreasing
38 Hg(0) levels in the NH over the last 20 years were partly caused by increased Hg(0) uptake by
39 vegetation due to increased terrestrial net primary production.

40

41 **Atmospheric mercury cycling**

42 Gaseous elemental mercury (Hg(0)) is the dominant form of natural and anthropogenic Hg
43 emissions and is transported globally through the atmosphere¹. Hg deposition to earth
44 surfaces occurs either by direct Hg(0) dry deposition or after oxidation to reactive Hg^{II}, which
45 is scavenged from the atmosphere by wet and dry deposition.² In addition to primary
46 anthropogenic emissions, Hg is also re-emitted as Hg(0) from [land and ocean](#) surfaces to the
47 atmosphere after reduction of Hg(II) pools in terrestrial and aquatic ecosystems².

48 The short-term balance between Hg emission, deposition and re-emission processes governs
49 diurnal and seasonal Hg(0) variations. For example, strong seasonal variations in atmospheric
50 Hg(0) concentration have been described for most background sites in the temperate
51 Northern Hemisphere (NH) with peaks in winter and minima in summer.³⁻⁶ Two widely
52 accepted hypotheses for the observed Hg(0) seasonality exist. The first hypothesis assumes
53 that Hg(0) variations are directly related to primary anthropogenic Hg(0) emissions from coal
54 combustion which are thought to be higher in winter due to higher energy demands for
55 heating.^{3,5} Global Hg models have not yet been able to test this [hypothesis](#), as current
56 anthropogenic Hg emission inventories have no seasonal resolution and are kept constant
57 throughout the year in models.⁷⁻⁹ [The second hypothesis suggests that atmospheric Hg\(0\)
58 oscillations are due to variations in](#) atmospheric Hg(0) oxidation rate and subsequent Hg(II)
59 deposition [which are both highest in summer in global Hg models.](#)⁷⁻¹⁰ Models incorporating
60 atmospheric Hg(0) oxidation [simulate equally-pronounced \(amplitudes of approx. 18%\)
61 seasonal Hg\(0\) variations in both Hemispheres. Hg\(0\) measurements at temperate sites in the
62 Southern Hemisphere \(SH\) however are constant with no seasonal oscillations.](#)^{6,11} Methane
63 [concentrations in the SH, for which atmospheric oxidation by OH radicals represents the most
64 important sink, do show strong seasonal variations.](#)¹² The absence of Hg(0) seasonality in the
65 [SH questions the relative importance of atmospheric Hg\(0\) oxidation.](#)

66 [Foliar uptake of atmospheric Hg\(0\) has long been recognized as a potentially important
67 pathway for atmospheric Hg deposition.](#)^{13,14} [The magnitude of terrestrial-atmosphere net
68 Hg\(0\) exchange \(foliar Hg\(0\) uptake - Hg\(0\) re-emission\) however has been debated and most
69 studies suggested a net emission of Hg\(0\).](#)¹⁵⁻¹⁸ [A recent review of Hg\(0\) flux measurements
70 over terrestrial surfaces concludes that there is no consensus whether terrestrial ecosystems
71 are a net sink or source for Hg\(0\) \(-513 to 1651 Mg a⁻¹, 25 to 75 percentile\)¹⁹. Most Hg\(0\) flux
72 studies however focused on soil re-emission, and thus neglected foliar Hg\(0\) exchange.¹⁹](#)

73 [There are several lines of evidence for net Hg\(0\) deposition to terrestrial ecosystems](#)
74 [that all invoke direct plant uptake of atmospheric Hg\(0\). First, foliage tissue Hg concentrations](#)
75 [increase continuously over the growing season, implying that Hg\(0\) is taken up by plants](#)
76 [through gas exchange and sequestered in leaf tissue.^{14,20} Structurally incorporated Hg in leaf](#)
77 [tissue derived from atmospheric Hg\(0\) therefore likely has a lower re-emission potential due](#)
78 [to photo-reduction,^{20,21} compared to Hg\(II\) deposited onto external leaf surfaces.²² Second,](#)
79 [global litterfall Hg deposition to soils - representing Hg\(0\) net uptake by foliage - is estimated](#)
80 [to be 1000-1200 Mg yr⁻¹ ^{17,23} representing half of primary anthropogenic emissions.²⁴ Third,](#)
81 [vegetation and soil Hg stable isotope signatures, which discern foliar Hg\(0\) uptake from Hg\(II\)](#)
82 [wet deposition, put firm constraints on plant Hg\(0\) uptake dominating \(50-80%\) Hg](#)
83 [deposition at terrestrial sites. ^{21,25-27}](#)

84 [Here, we investigate the role of Hg\(0\) uptake by plants in modulating global](#)
85 [atmospheric Hg\(0\) seasonality. We compare the atmospheric dynamics of Hg\(0\) with those of](#)
86 [CO₂ - a tracer for gas exchange by plants - for which seasonal oscillations in the Northern](#)
87 [Hemisphere \(NH\) are attributed to the balance of CO₂ assimilation and respiration during the](#)
88 [year.²⁸⁻³⁰](#)

90 **Correlation of Hg(0) and CO₂ seasonality**

91 We [investigated](#) multi-year time series of Hg(0) [and CO₂](#) at seven [NH and SH background sites](#)
92 [\(mean Hg\(0\) < 2 ng/m³\) and Hg\(0\) at an additional 43 global sites. Multi-year Hg\(0\) and CO₂](#)
93 [observations of the five NH background sites, normalized for linear long-term trends, show](#)
94 [seasonal variations with highest concentrations in winter and minima in summer and fall \(**Fig.**](#)
95 [1, S1, S3 and S6\).](#) For all NH sites, we found a significant positive correlation between Hg(0)
96 and CO₂ (Pearson's product-moment correlation of monthly means, normalized for long-term
97 trends, Birkenes, Norway (r=0.37, p<0.005), Schauinsland, Germany (r=0.50, p<0.001), Mace
98 Head, Ireland (r=0.40, p<0.001), Pallas, Finland (r=0.54, p<0.001) and Egbert, Canada (r=0.53,
99 p<0.001)). We investigated two coastal SH sites, Cape Point (South Africa) and Amsterdam
100 Island (France), with the latter surrounded by 3000 km of ocean on all sides. Cape Point
101 showed a weaker (r=0.30, p<0.01) correlation and no clear seasonal variation and Amsterdam
102 Island did not show a significant correlation between the two trace gases. Both SH sites show
103 small day-to-day variability, small diurnal variation and small seasonal amplitudes in CO₂, (<3
104 ppm, <1%) and Hg(0) (<0.1 ng/m³, <10%) (**Fig. 1**). This suggests low anthropogenic

105 contributions (i.e., no pollution plumes) as well as atmosphere-ocean exchange of Hg(0) that
106 is relatively constant throughout the day and seasons.

107 We find that the interhemispheric gradient in Hg(0) concentrations ($\Delta\text{Hg}_{45^\circ\text{N}-45^\circ\text{S}}$) is
108 largest in February (0.66 ng/m³) and smallest in September (0.43 ng/m³) with a yearly
109 average of 0.54 ng/m³ (Fig 2A). Two thirds of the primary anthropogenic emissions are to the
110 NH,³¹ which is the main explanation for the higher NH Hg(0) concentrations compared to the
111 SH.⁶

112 The seasonal amplitudes of Hg(0) measurements at temperate NH background sites
113 are much larger, averaging 0.39 ng/m³ (equivalent to 25% of the average annual Hg(0)
114 concentration of 1.5-1.7 ng/m³),⁶ and were 6 times larger than the corresponding CO₂
115 amplitudes (4%). Excluding Polar sites (see below), we find a positive correlation ($r=0.88$,
116 $p<0.01$, Pearson's product-moment correlation) between seasonal Hg(0) (Sep-Feb) and
117 CO₂ (Aug-Feb) amplitudes (**Fig. 2C**). Amplitudes of seasonal CO₂ oscillations are largest at high
118 NH latitudes and decrease towards the south with minimal oscillations in the SH, which has
119 been related to varying CO₂ exchange during the vegetation period in the NH, and fewer land
120 masses in the SH.²⁸⁻³⁰ We find a significant correlation between the seasonal Hg(0) amplitudes
121 of all background sites (n=38) and latitude ($R^2 = 0.42$, $p<0.001$, Figure 2B). However, Hg(0)
122 amplitudes are confounded by additional factors. For example, we explain a lower amplitude
123 of Egbert (Ca) by relatively high contributions of polluted air masses throughout the year
124 (Figure S12). Comparing all investigated sites, we find significantly lower Hg(0) amplitudes at
125 coastal sites compared to inland background sites, both for Europe (0.25 ng/m³ vs. 0.39
126 ng/m³, $n=13$, $p<0.05$, 2-sided t-test) and North America (0.27 ng/m³ vs. 0.41 ng/m³, $n=18$,
127 $p<0.05$, 2-sided t-test). A special case is a different seasonality of Hg(0) in the Arctic, which is
128 typically characterized by spring minima due to sea-salt induced atmospheric mercury
129 depletion events (AMDEs) and an unusual summertime maximum.³² Observations from SH
130 terrestrial background sites are few, yet the first measurements in Bariloche, Patagonia (Ar)
131 revealed seasonal Hg(0) oscillations with minima in SH summer and autumn.³³

132 If seasonal variations in atmospheric Hg(0) were mainly due to oxidation rates and
133 ocean re-emission,⁷⁻⁹ one would expect higher Hg(0) amplitudes at coastal sites and in the SH,
134 which is contrary to the patterns described above. On an ecosystem scale, studies have
135 reported enhanced Hg deposition fluxes during vegetation growth periods^{21,34} and lower local
136 atmospheric Hg(0) levels over vegetated surfaces.^{35,26,36} The fact that the highest seasonal
137 amplitudes are observed at inland terrestrial background sites — along with similar patterns

138 observed for CO₂ — argues for a dominant role **of** vegetation Hg(0) uptake **in** modulating
139 global atmospheric Hg(0) concentrations. The NH is not only more populated but also has a
140 much larger land mass (68%) compared to the SH (32%), resulting in higher plant Hg(0)
141 uptake which may explain the weaker interhemispheric gradient during the NH growing
142 season.

143 Most NH sites reveal a remarkable 1-month offset in Hg(0) summertime minima (Sep)
144 compared to CO₂ (Aug) (Fig. S3, S6). Opposite oscillation in diurnal Hg(0) and CO₂
145 concentrations (Fig. S4) and later seasonal minima of Hg(0) relative to CO₂ suggests that Hg(0)
146 uptake by plants is continuing during periods of net CO₂ respiration during nighttime and in
147 fall, when the ecosystem exchange of CO₂ turns from net assimilation to net respiration.³⁷
148 Hg(0) thus follows a similar behavior as carbonyl sulfide (COS), for which seasonal patterns
149 are coupled to CO₂, whereas diurnal variations are decoupled.^{38,39} This suggests that Hg(0)
150 plant uptake is controlled by plant gas exchange as represented by CO₂ here. Foliar Hg(0) re-
151 emission however is decoupled from CO₂ respiration and is likely limited. This is reflected in
152 increasing Hg concentrations in foliage over the growing season.^{13,20,35} The observed Hg(0)
153 concentration minimum in September may be amplified by a shallower planetary boundary
154 layer (PBL) in autumn leading to a reduced mixing with background air.³⁰

156 Variations in Hg(0) vs. vegetation Index

157 We compared monthly Hg(0) concentrations with the satellite-based Normalized Differenced
158 Vegetation Index (NDVI), representing the vegetation photosynthetic activity and thus a proxy
159 for plant gas exchange, at the location of the 44 NH monitoring sites (0.1° resolution). We **find**
160 a significant correlation between monthly Hg(0) and NDVI for for 22 of the 44 investigated
161 sites. For the NH background sites where Hg(0) and CO₂ was measured, NDVI significantly
162 correlated with both trace gases (Fig. 3, Table S1), with the exception of Hg(0) at Schauinsland,
163 De. **For example, at** Pallas NDVI was able to explain 69 % and 82% of seasonal Hg(0) and CO₂
164 variation, respectively. The correlation was strongest at background sites in Canada ($R^2 > 0.75$
165 for e.g. Burnt Island, or Kejimikijuk, Figure S8, Table S2), Scandinavia ($R^2 > 0.6$ for e.g. Andoya,
166 Bredkåle, or Vavihill, Fig. S9, Table S3) or for some Mountain sites in China ($R^2 \approx 0.5$ for e.g. Mt.
167 Damai, Mt. Leigong, Mt. Chanbai, Fig. S10, Table S4). For urban sites in North America (e.g. Salt
168 Lake City, Houston or Detroit, Fig. S7, Table S2) or Asia (Beijing, Nanjing, Fig. S10, Table S4)
169 we did not find a significant correlation between Hg(0) and NDVI (red points in **Fig. S8**),
170 indicating that local anthropogenic emissions dominate Hg(0) levels at urban sites. The

171 relatively strong correlation between Hg(0) concentrations and [NDVI](#) at regional background
172 sites suggests that vegetation uptake is responsible for Hg(0) depletion in summer months,
173 when vegetation [photosynthetic activity and thus gas exchange of plants](#) is strongest (high
174 NDVI). [Note that NDVI is able to describe a higher degree of the Hg\(0\) variance compared to](#)
175 [CO₂ concentrations which we relate to the fact that CO₂ respiration from terrestrial](#)
176 [ecosystems is decoupled from Hg\(0\) re-emission.](#) We therefore suggest that NDVI satellite
177 data might be a robust parameter to estimate terrestrial Hg(0) uptake fluxes on a spatial and
178 temporal scale. Field studies comparing Hg(0) fluxes with vegetation cover are needed in
179 order to validate this approach. [A recent study found a strong correlation of soil Hg pools with](#)
180 [NDVI across western North America,⁴⁰ suggesting a long-term control of the plant Hg\(0\) pump](#)
181 [on soil Hg pools, which is strongly supported by Hg stable isotope data.^{21,25-27}](#)

183 [Seasonality of Hg\(0\) emissions approximated by coal consumption statistics](#)

184 [Coal combustion represents a major primary Hg\(0\) emission source to the atmosphere.³¹ We](#)
185 [estimated the seasonal variation of anthropogenic Hg emissions \(Table S4\) by using monthly](#)
186 [coal consumption data reported for the United States, Europe \(EU27\) and China. The](#)
187 [estimates show a peak in Hg emission during the cold winter months attributed to increased](#)
188 [energy consumption for heating \(Fig. 4, Table S4\). In North America, there is a second peak in](#)
189 [coal consumption and thus Hg\(0\) emission \(Fig. 4, Table S4\) during hot summer months that](#)
190 [exceeds the winter peak and is attributed to high energy consumption for air-conditioning of](#)
191 [buildings.⁴¹ Hg\(0\) concentrations measured at three urban monitoring sites in North America](#)
192 [\(Salt Lake City, Houston and Detroit\)^{42,43} closely follow our seasonal emission estimates](#)
193 [including a noticeable summertime Hg\(0\) increase \(Fig. 4\). Regional background sites in](#)
194 [North America however show a consistent summertime minimum \(Fig. 4\), which cannot be](#)
195 [explained by anthropogenic Hg emissions, implying that other processes drive the lower](#)
196 [Hg\(0\) levels in summer. Inverse modeling, that optimized terrestrial-atmosphere Hg\(0\)](#)
197 [exchange fluxes to fit observational Hg\(0\) data, suggested enhanced summertime net Hg\(0\)](#)
198 [deposition over eastern North America⁴⁴ supporting the strong role of vegetation as a pump](#)
199 [for Hg\(0\). In Europe and Asia, the anthropogenic Hg\(0\) emission estimates based on coal](#)
200 [consumption data are highest in winter \(Table S4\) with no summertime increase and a](#)
201 [seasonality similar to Hg\(0\) observations \(Fig. 4\). Higher Hg\(0\) concentrations during](#)
202 [summer months at urban sites in Asia have been attributed to enhanced re-volatilization of](#)
203 [deposited Hg.⁴⁵ For the Waldhof site in central Europe, regional models have found a strong](#)

204 [correlation between emission and particulate mercury but not for Hg\(0\).](#)⁴⁶ We therefore
205 [conclude that variations in primary anthropogenic emissions alone cannot explain the](#)
206 [observed seasonal Hg\(0\) variation in the NH.](#)

208 **Implications for global Hg cycling**

209 Hg(0) concentrations measured in the PBL at terrestrial background sites reflect both
210 deposition and emission processes. Therefore observed Hg(0) oscillations can be considered
211 as variations in net exchange (vegetation uptake – [soil/vegetation](#) re-emission). The strong
212 depletion of atmospheric Hg(0) observed at terrestrial background sites in summer, despite
213 highest solar radiation and therefore potential photo-reductive re-emission, suggests that
214 terrestrial ecosystems are net sinks for Hg(0). We therefore conclude that vegetation uptake
215 of Hg(0) is large and dominates over other terrestrial emission and deposition processes at a
216 regional and global scale. Considering an atmospheric Hg(0) pool of [4800](#) Mg,⁴⁷ the 20%
217 amplitude of seasonal Hg(0) oscillations indicate that terrestrial ecosystems are drawing
218 approximately [1000](#) Mg a⁻¹ of Hg(0) from the atmosphere [via](#) the vegetation pump, which is
219 in agreement with foliage/litterfall estimates^{17,23}. This suggests that approximately half of the
220 [annual](#) primary anthropogenic emissions are assimilated by terrestrial vegetation, where it [is](#)
221 efficiently retained against re-emission to the atmosphere, though susceptible to transfer [via](#)
222 [soils](#) to [continental and coastal](#) aquatic ecosystems.²¹

223 We suggest that the vegetation pump controls to a large extent diurnal and seasonal
224 cycling of atmospheric Hg(0) in the terrestrial PBL, which has large implications for global Hg
225 cycling and interpreting and forecasting long-term trends. The absence of Hg(0) seasonality
226 observed in the SH seems inconsistent with our current understanding of fast Hg(0) oxidation
227 in the atmosphere^{7,9,47}. A dominant role of plants in NH Hg(0) seasonality [may](#) imply that
228 atmospheric [Hg\(0\)](#) oxidation [is less important](#) than currently assumed. Average Hg(0) levels
229 in the NH have decreased by 20-40 % between 1990 and 2010 and were recently attributed
230 to a cumulative 30% (≈ 600 Mg a⁻¹) decrease in global primary anthropogenic Hg(0)
231 emissions.²⁴ Over the same time period, net primary production (NPP) is thought to have
232 increased in the NH by 0.13-0.15 Pg C a⁻¹.^{48,49} Assuming median foliar Hg concentrations of 24
233 ng g⁻¹,¹⁷ we estimate that Hg(0) deposition by NH vegetation uptake today has increased by
234 ≈ 140 Mg a⁻¹ relative to 1990. We therefore suggest that a significant fraction of the observed
235 decrease in NH atmospheric Hg(0) concentrations [s](#) resulted from increased vegetation uptake
236 of Hg(0). Recent estimates of gross primary production (GPP), based on atmospheric COS

237 cycling, suggested even larger increases in GPP by 31% during the twentieth century⁵⁰, that
238 may have strongly counterbalanced estimated increases in Hg(0) emissions. Regional
239 reductions in vegetation cover by deforestation⁵¹ or droughts⁴⁹ on the other hand would slow
240 down or cease the vegetation Hg(0) pump, potentially resulting in higher atmospheric Hg(0)
241 concentrations. We therefore emphasize the need to incorporate seasonal and spatial
242 variability in vegetation uptake of Hg(0) into global Hg models to better assess its impact on
243 regional and global Hg cycling. Trends in vegetation cover should be incorporated in models
244 reconstructing past and predicting future Hg(0) levels. The effects of Hg(0) uptake by
245 vegetation related to climate change and land-use change should be considered when
246 discussing mitigation strategies to reduce human Hg exposure. Finally, the paradigm shift we
247 propose regarding the dominant global atmospheric Hg deposition pathway demands revised
248 Hg deposition monitoring strategies by environmental agencies.

249

250 **Data, Materials & Methods**

251 Atmospheric CO₂ and Hg(0) measurements. The focus of the present analysis is on
252 atmospheric monitoring sites that simultaneously measured Hg(0) and CO₂ concentrations.
253 As atmospheric oxidized mercury species account for a minor fraction of total atmospheric Hg
254 (Hg_{tot}) in the PBL, no difference was made between Hg(0) and Hg_{tot} and all data are reported
255 as Hg(0). High resolution QA/QC controlled Hg(0) and CO₂ data were obtained from global air
256 monitoring databases or directly from the responsible site scientists. The European data are
257 harmonised and quality assured following the EMEP program and monitoring strategy, see
258 [Tørseth, et al. 52](#) Hg(0) concentrations at all sites except Pallas were measured at high
259 frequency (5-15 min) using Tekran Continuous Mercury Vapor Analyzers 2537 and reported
260 as 1h averages. For Cape Point, SA and Amsterdam Island, Fr all Hg(0) measurements were
261 made in compliance with the GMOS quality assurance protocol (www.gmos.eu). Hg(0)
262 concentrations at Pallas were measured using a semi-automatic sampling on Au-traps (quartz
263 glass coated with gold) at an air sampling rate of 300 ml/min during 24h periods, two days
264 per week. The samples were analyzed on a Tekran 2500 CVAFS Mercury Detector at the IVL
265 laboratory in Gothenburg. CO₂ at Birkenes was measured with Cavity Ringdown spectroscopy
266 (CRDS), Picarro G1301 with 5 s resolution, average to 1 h if data coverage was more than 75%.
267 The data are calibrated against WMO- Global Atmospheric Watch standards at EMPA. CO₂ data
268 from Amsterdam Island, Fr was measured using CRDS and from Egbert, Ca and Cape Point, SA
269 using Non-Dispersive Infrared (NDIR) sensors. CO₂ data from Schauinsland, De were
270 measured by IR-Absorption. All data from Birkenes, No, Amsterdam Island, Fr, Egbert, Ca,
271 Cape Point, SA and Schauinsland, De were measured continuously and reported as hourly
272 averages. CO₂ data from Pallas, Fi and Mace Head, Ir were measured on an event base with
273 flasks and analysed using NDIR.⁵³ Monthly average Hg(0) data were obtained from [4,42,43,45,54-](#)
274 [58,6,59-61](#)

275 Normal Differenced Vegetation Index, (NDVI) were obtained from from the NASA Earth
276 Observations (NEO) platform at 16 days and 0.1° resolution. NDVI data are derived from
277 Moderate Resolution Imaging Spectroradiometer (MODIS) images aboard the NASA's Terra
278 and Aqua satellites.

279 **Statistics**

280 The Co-variance between Hg(0) and CO₂ was assessed using Pearson's product moment
281 correlation coefficient with the cor.test function of the stats (v3.1.3) package and long-term
282 trends of time series as well as linear regression's of Hg(0) and CO₂ with NDVI were assessed

283 with the lm function of R.⁶² Students t-tests were performed with Microsoft Excel assuming
284 uneven variances.

285 Seasonality of anthropogenic Hg emission. The monthly Hg emission for the different regions
286 (F(Hg)_{month}) in Mg Hg/month was calculated according to Formula 1:

$$287 \quad F(\text{Hg})_{\text{month}} = f_{\text{coal}} * f_{\text{consumption}} * F\text{Hg}_{\text{year}} + (1 - f_{\text{coal}}) * \frac{1}{12} * F\text{Hg}_{\text{year}} \quad (1)$$

288 where f_{coal} and $F(\text{Hg})_{\text{year}}$ corresponds to the fraction of Hg emission based on coal combustion
289 and the total yearly Hg emission for each investigated country, based on the AMAP/UNEP
290 2010 mercury emission estimates⁶³. $f_{\text{consumption}}$ corresponds to the monthly fraction of the
291 yearly coal consumption. For the USA and Europe (27 EU countries) monthly coal
292 consumption data are reported by governmental agencies. Seasonality of Hg emission from
293 China was derived from the monthly power generation data between 2005 and 2010 reported
294 by Liu et al. 2015⁶⁴. The monthly cement production in China was considered constant
295 throughout the year⁶⁵.

296 Normalization of monthly Hg concentration and emission data. In Figure 1 monthly Hg
297 concentrations and emission fluxes are normalized to yearly averages for comparability. Hg
298 concentrations are normalized as follows: $C(\text{Hg0})_{\text{month}x} / C(\text{Hg0})_{\text{year}}$, were $C(\text{Hg0})_{\text{month}x}$
299 represents the multiyear monthly average of month x and $C(\text{Hg0})_{\text{year}}$ the yearly average.

300 Computation of Seasonal Amplitudes. The absolute seasonal amplitudes are calculated from
301 the multiyear monthly averages as follows: $(\mu\text{M}_{\text{max}} - \mu\text{M}_{\text{min}})$, were μM_{max} and μM_{min} represent
302 the averages of the months with NH maximum (February for both Hg(0) and CO₂) and
303 minimum (August for CO₂ and September for Hg(0)) concentration (Table S6-S10),
304 respectively. Relative amplitudes are calculated as follows: $(\mu\text{M}_{\text{max}} - \mu\text{M}_{\text{min}}) / \mu\text{Y}$, were μY
305 represents the yearly average.

306

307 **Data availability**

308 Most European Hg(0) data used are associated to EMEP, and publically available through the
309 EMEP data base, EBAS: <http://ebas.nilu.no>. CO2 concentrations were obtained from the ESRL
310 NOAA Global Monitoring Division database
311 ([ftp://aftp.cmdl.noaa.gov/data/trace_gases/co2/flask/surface/co2_mhd_surface-](ftp://aftp.cmdl.noaa.gov/data/trace_gases/co2/flask/surface/co2_mhd_surface-flask_1_ccgg_event.txt)
312 [flask_1_ccgg_event.txt](ftp://aftp.cmdl.noaa.gov/data/trace_gases/co2/flask/surface/co2_mhd_surface-flask_1_ccgg_event.txt),
313 [ftp://aftp.cmdl.noaa.gov/data/trace_gases/co2/flask/surface/co2_pal_surface-](ftp://aftp.cmdl.noaa.gov/data/trace_gases/co2/flask/surface/co2_pal_surface-flask_1_ccgg_event.txt)
314 [flask_1_ccgg_event.txt](ftp://aftp.cmdl.noaa.gov/data/trace_gases/co2/flask/surface/co2_pal_surface-flask_1_ccgg_event.txt)) and the World Data Center for Greenhouse Gases WDCGG
315 (<http://ds.data.jma.go.jp/gmd/wdcgg/>). Canadian Hg(0) concentrations were obtained from

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316 the National Atmospheric Chemistry (NAtChem) Database and Analysis Facility of
317 Environment Canada (www.ec.gc.ca/natchem).
318 For the USA and Europe (27 EU countries) monthly coal consumption are reported by the U.S.
319 Energy Information Administration (<http://www.eia.gov/totalenergy>, downloaded
320 25.5.2016) and Eurostat ([http://ec.europa.eu/eurostat/statistics-
321 explained/index.php/Coal_consumption_statistics](http://ec.europa.eu/eurostat/statistics-explained/index.php/Coal_consumption_statistics), downloaded 15.2.2016), respectively.
322 NDVI data was obtained from the NASA Earth Observations (NEO) platform
323 (<http://neo.sci.gsfc.nasa.gov>)

324
325
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334
335 **Author Contribution**
336 MJ initiated the project, performed the data analysis and wrote the manuscript together with
337 JES and contributions from DO, JB and AD. JB, RE, CML, KA, IW, KK, DW, LM, CL, TM, MR, OM,
338 and AD provided data. All [Authors](#) read and commented on the manuscript.

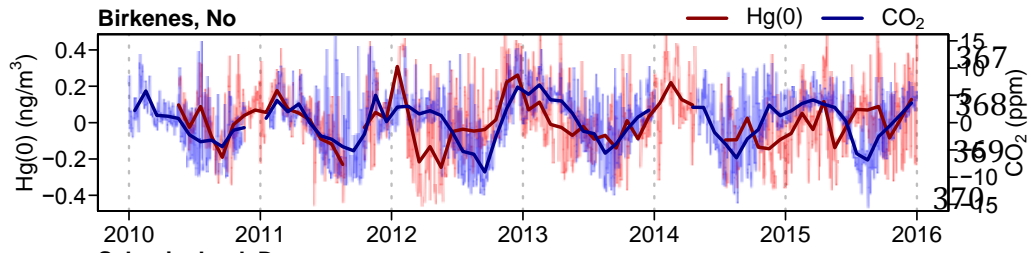
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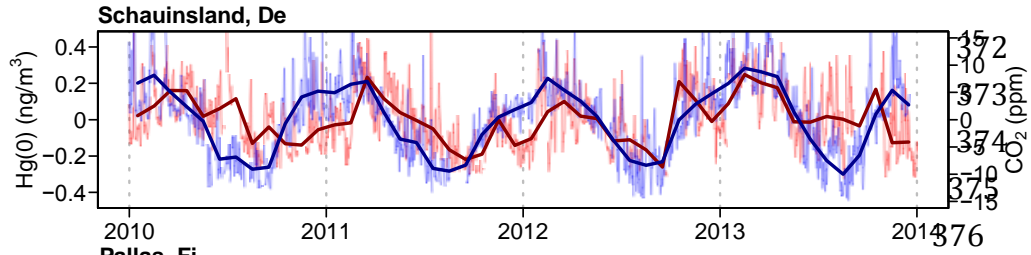
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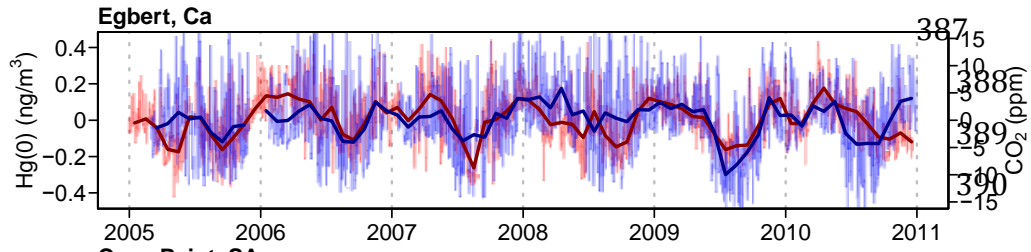
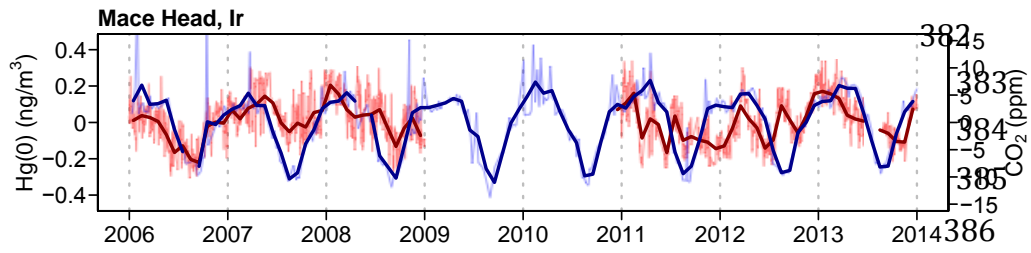
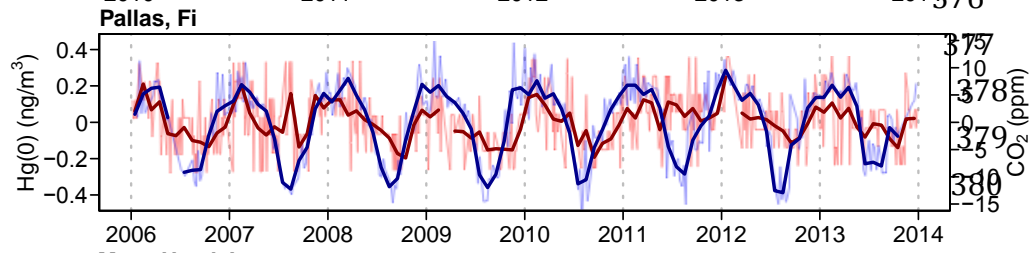
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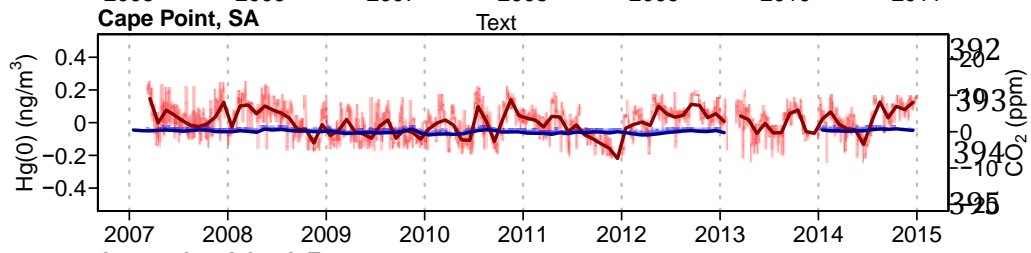
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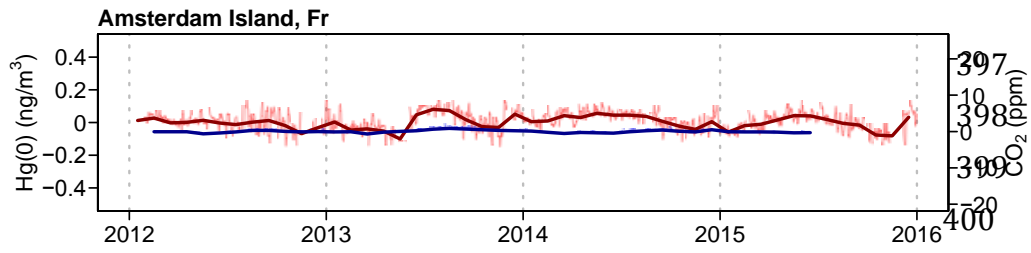
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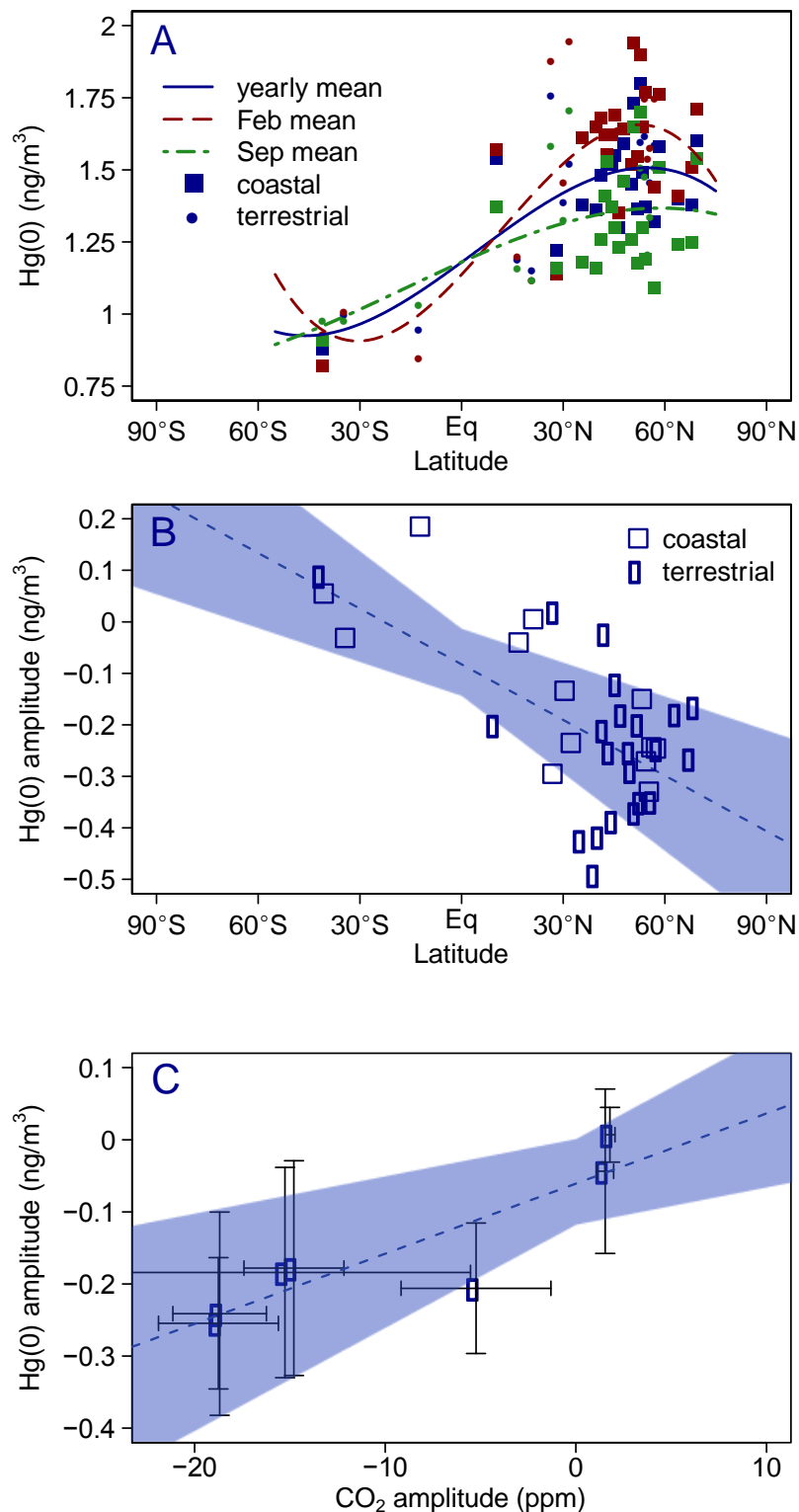
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Fig. 1: Seasonal oscillations of atmospheric Hg(0) (red) and CO₂ (blue) concentration at several monitoring sites around the globe. Fine lines represent daily and bold lines represent monthly

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404 [anomalies normalized for long-term linear trends. All data meeting QA/QC criteria of the](#)
405 [respective site are presented.](#)



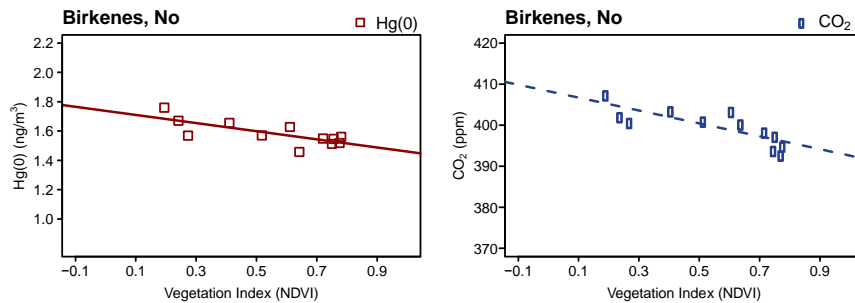
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410 **Fig. 2: [A: Interhemispherical Gradient of Hg\(0\) at background sites \(<2 ng/m³\). The lines](#)**
411 **[represent polynomial fits of the measurements for the yearly mean and the monthly means of](#)**
412 **[February and September. \[B: Seasonal amplitude of Hg\\(0\\) \\(September mean - February mean\\)\]\(#\)](#)**
413 **[as a function of latitude for coastal \(squares\) and terrestrial \(circles\) sites. The dashed line](#)**

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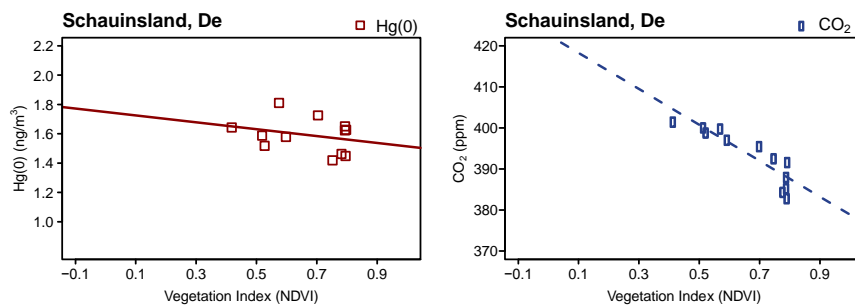
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414 [represents the linear fit of the data and the shaded area the 95% confidence area. C: Seasonal](#)
415 [amplitudes of Hg\(0\) \(Sep-Feb\) and CO₂ \(Aug-Feb\) variation measured at 7 atmospheric sites.](#)
416 [The error bars represent the interannual variability \(1σ\). The dashed line represents the linear](#)
417 [fit of the data and the shaded area the 95% confidence area.](#)
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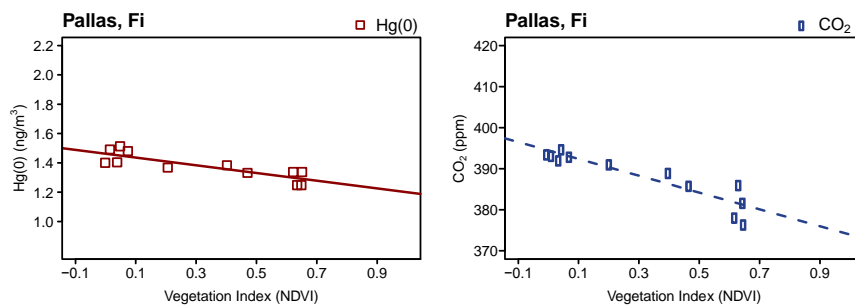
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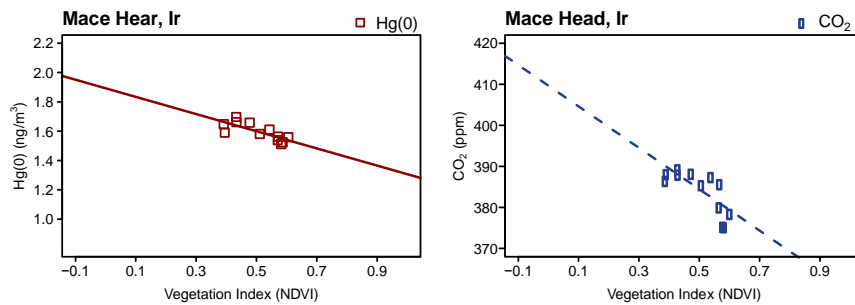
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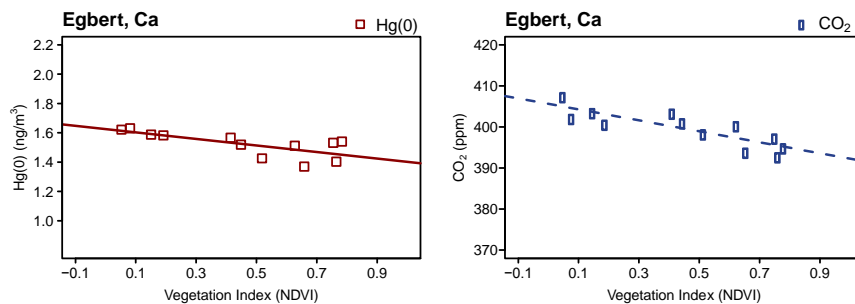
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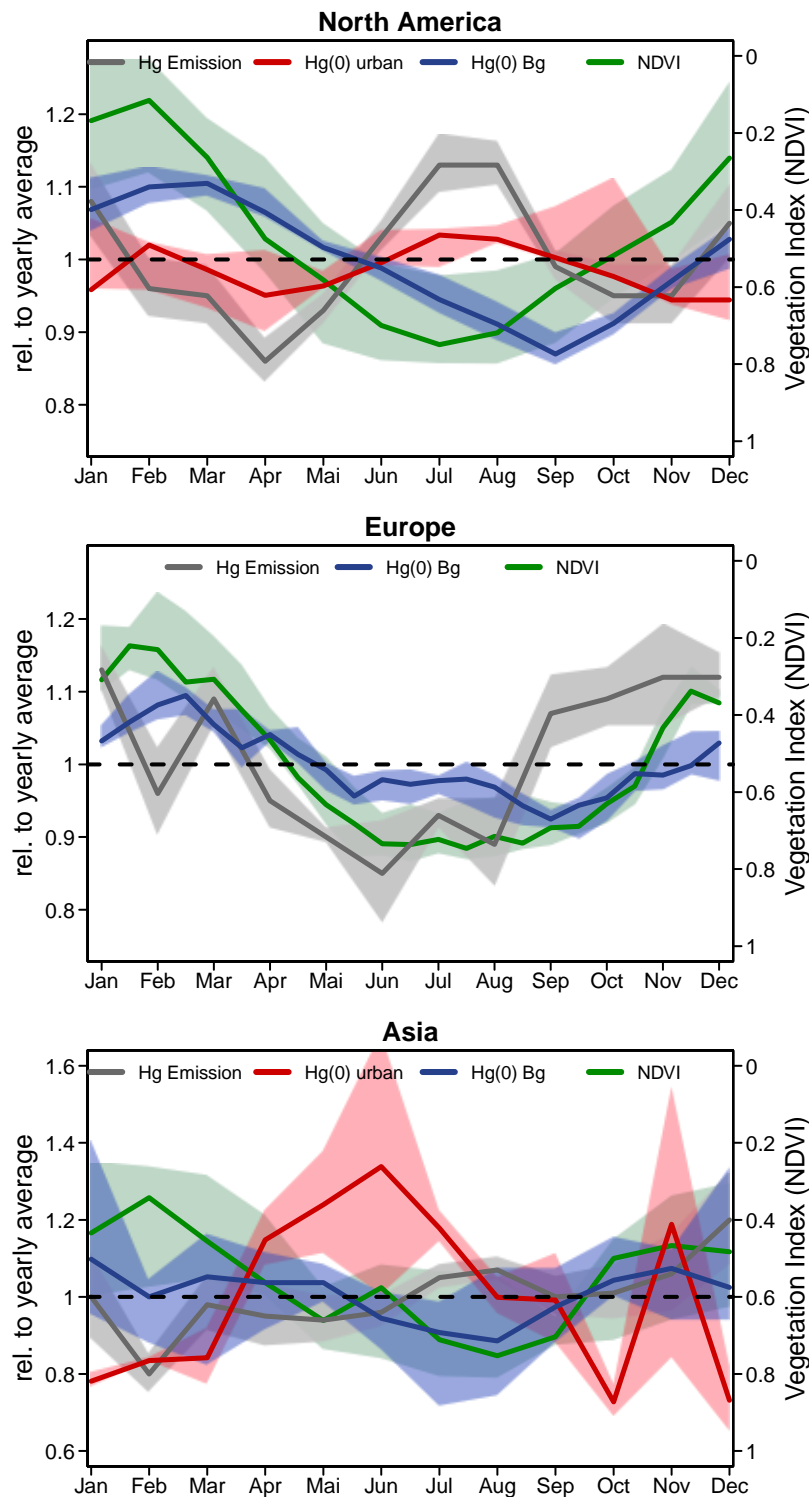


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Figure 2: Linear correlation between multi-year averages of monthly Hg(0) concentration and NDVI (left) and CO₂ concentration and NDVI (right) at five NH monitoring stations.



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Fig. 4: Seasonal variation of Hg emissions in gray, atmospheric Hg(0) concentration at terrestrial background sites (Hg(0) Bg) in blue and at urban sites (Hg(0) urban) in red and Vegetation index (NDVI) in green. Monthly Hg emission and Hg(0) concentrations are reported

432 relative to yearly average. The bold lines represent averages and the shaded areas represent
433 25th to 75th percentiles.

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