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## Effect of demand-controlled ventilation strategies on indoor air pollutants in a classroom: A Norwegian case study

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Abstract. The choice of the minimum ventilation rate (V<sub>min</sub>) in a demand-controlled ventilation strategy can influence energy demand but also introduce outdoor air pollutants. The latter may have direct health effects, as well as affect indoor chemical reactions. In this paper, we evaluate the effect of ventilation rates and operation hours on the level of  $CO_2$ , nitrogen dioxide (NO<sub>2</sub>), and ozone  $(O_3)$  in a classroom during normal use. We compared the baseline ventilation scenario (S0) with a  $V_{min}$  of 430 m<sup>3</sup>/h with S1;  $V_{min}$  of 150 m<sup>3</sup>/h for normal ventilation operation time (6:30-17:00) and continuous ventilation for 24h (S2). We found that S1 with reduced V<sub>min</sub> would lower the ozone concentration by 35% during the hours before occupancy compared to S0. Moreover, continuous ventilation during night time with a low V<sub>min</sub> resulted in almost as high O<sub>3</sub> concentrations as the baseline ventilation scenario. As O<sub>3</sub> reacts easily with certain VOCs to produce secondary organic aerosols, the level of V<sub>min</sub> and the ventilation duration would impact the indoor air quality upon entering the classroom.

#### 1. Introduction

For buildings with varying occupancy or pollution load, especially in offices and schools where rooms are unoccupied for most of the operational time, demand-controlled ventilation (DCV) can significantly improve energy performance compared to constant air volume (CAV) ventilation [1–3]. Seppänen [4] summarized how different (ventilation) strategies could reduce energy consumption and improve indoor air quality. Among them is demand-controlled ventilation, which may enhance air quality and yield up to a 50% reduction in energy usage for large spaces. Thus, efforts have been made to design DCV systems that provide recommendable air quality with little input in energy consumption.

The ventilation airflow rate in modern DCV systems is controlled between the predefined minimum  $(V_{min})$  and maximum  $(V_{max})$  limit values, based on the signal from one or more room sensors, typically CO2 and/or measured air temperature. The Vmin and Vmax limit values are set to account for, for example, pollutant load from materials, room size, or the maximum pollutant load from occupants. The choice of  $V_{min}$  has an obvious impact on energy use, but there are few studies on the impact on indoor air quality (IAQ) [5-7]. Previous studies performed in unoccupied classrooms have shown no significant effect on perceived air quality (PAQ) by increasing V<sub>min</sub> above 1.0 l/s/m<sup>2</sup> for classrooms with no additional pollutant sources [6,7].

A possible rationale for choosing the ventilation rate is to maintain a constant IAQ based on sensory pollution strength (olf) suggested by Fanger [8]. This implies that V<sub>min</sub> is set to achieve an acceptable olfactory level accounting for the total pollutant load from materials and that ventilation is increased

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above  $V_{min}$  according to pollutants from occupants in the room. For unoccupied rooms, the minimum ventilation demand varies typically from 0.7 to over 2.0 l/s/m<sup>2</sup> in Norway.  $V_{min}$  is often set to the upper end of this range due to the risk of high-emitting furniture. While this strategy may be a robust choice to account for unexpected pollutant sources, it increases energy demand and introduces outdoor pollutants at a higher rate. Some of these may have direct health effects, as well as affect indoor chemical reactions [9,10].

The objective of the study was to examine the interplay between ventilation strategy and outdoor as well as stationary and transient indoor pollutant sources. In this paper, we evaluate the effect of ventilation rates and operation hours on the level of  $CO_2$  concentration, nitrogen dioxide (NO<sub>2</sub>), and ozone (O<sub>3</sub>) in a classroom during normal use by 6-7-year-old pupils.

### 2. Methods

### 2.1. Study design

The measurements took place in a classroom at a primary school located 500 m southwest of a highway north of central Oslo, Norway. The three-story school building was completed in August 2016 after the Norwegian passive house standard [11]. The classrooms are similar in size, with an average floor area of 60 m<sup>2</sup> (59.7) and a height of 2.8 m, and have similar furnishings. The selected classroom is situated on the second floor and is occupied by first graders (aged 6-7 years).

A full class consisted of 24 pupils and two teachers. The first graders had no fixed daily schedules. Oftentimes the class would be divided into groups, so the classroom was frequently not fully occupied. A typical school day started at 8:30 and finished at approximately 14:45. The teaching sessions with a full class were few and varied in time and duration. Most of the time, the pupils were either outside playing or split into groups to participate in various activities in other classrooms. Nevertheless, there were two eating breaks daily when the entire class was gathered, one at 10:30 and one at 14:15. Due to the high frequency of change in pupils during classroom time, security and GDPR, automatic occupancy detection (e.g., sensor counters, cameras or survey) with ground truth were not established but was performed manually.

The school has balanced supply and exhaust mechanical ventilation with a cooling battery, rotary heat exchanger, and water heating battery. If there is a cooling demand for supply air and a lower extract temperature than the outdoor temperature, the heat recovery unit will act as a cooling recovery unit. The ventilation airflow rates in all classrooms were demand-controlled by CO<sub>2</sub>- and temperature signals from a room sensor. When CO<sub>2</sub> exceeds 500 ppm and/or room temperature exceeds 23 °C, ventilation rates are increased above  $V_{min}$  by opening dampers. An optimizer system adjusts the fan power to achieve demanded airflow rates with as low duct pressure as possible [12].

The ventilation system had glass fiber bag filters of class F7 type Standard-Flo XLT 7 with a minimum efficiency of 41% for particles  $<1\mu$ m diameter, but no carbon filtering to reduce gaseous compounds. The bag filters were replaced in July 2017. The normal ventilation operation time was between 6:00 - 17:00 during the weekdays, and the ventilation was turned off during the weekend.

	Ventilation operation time	$V_{min}$ $(m^3/h)/(l/s/m^2)$	V <sub>max</sub> (m <sup>3</sup> /h)	CO2 setpoint (ppm)	Temperature setpoint (°C)
Baseline S0	06:30-17:00	430 / 2.0	1060	500	25
<b>S</b> 1	06:30-17:00	150 / 0.7	1060	1000	25
S2	00:00-24:00	150 / 0.7	1060	1000	25

Table 1. Overview of the three ventilation scenarios for the weekdays.

An overview of the three ventilation scenarios and the relevant setpoints is shown in Table 1. For the three-week-long measurement campaign, between 28 May and 18 June 2018, the ventilation operation time was manually adjusted to 6:30 - 17:00 during weekdays and 8:00 - 17:00 during weekends. The ventilation was turned off during the third weekend. As the measurement campaign took place during

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the year's warmest month, a higher temperature setpoint was chosen to prevent the ventilation system from operating with maximum airflow during unoccupied hours. Due to time constraints and the availability of the study site, it was not possible to repeat the measurement campaign for other periods.

The concentration of ozone  $(O_3)$  in the classroom was monitored using a UV absorption analyzer (Thermo Model 49C  $O_3$  Analyzer). Nitrogen oxides (NOx, consisting of NO<sub>2</sub> and NO) were measured in the classroom using a chemiluminescence analyzer (Teledyne API 200A). Both parameters were logged with a 1-minute time resolution.  $O_3$  was also measured in the supply air every 15<sup>th</sup> minutes.

The building management system (Desigo CC, Siemens Switzerland Ltd) provided measurement data on carbon dioxide  $(CO_2)$ , temperature, and ventilation airflow rates. Log data from the building management system (BMS) is event-based, meaning a value is only provided if there is a change from the previous value. For comparison purposes, we also placed two Rotronic CP 11 (Rotronic AG, Bassersdorf, Switzerland) on top of a cupboard inside the classroom, approximately 30 cm above the room sensor, to log room temperature, relative humidity (RH), and CO<sub>2</sub> concentration with a 1-minute and 5-minute time resolution. Three Tinytag Plus 2 (Gemini Data Loggers, UK) were placed in different locations in the classroom; next to the Rotronics, in the extract air duct, and on the windowsill. These were also logged at 5-minute intervals. In addition, hourly mean weather statistics including outdoor air temperature and RH were also collected from the local weather monitoring station (via seklima.met.no, Norwegian Centre for climate services). Advanced measurements of volatile organic compounds (VOCs) and size and number concentration of particles with diameters from 10nm to 350 nm were also performed, but not analysed in this paper. The sampling campaign with an overview of the different ventilation scenarios is shown in Table 2. The different ventilation scenarios were manually adjusted by changing the setpoints in the centralized BMS, while the airflow rates (V<sub>min</sub>) were adjusted using a handheld ZTH (Belimo) that connects directly to the VAV-controller.

Table 2. Overview of the different ventilation scenarios during the sampling campaign.

Week	Monday	Tuesday	Wednesday	Thursday	Friday	Saturday	Sunday
22	S0	SO	<b>S</b> 1	<b>S</b> 1	<b>S</b> 1	<b>S</b> 1	<b>S</b> 0
23	<b>S</b> 2	<b>S</b> 2	<b>S</b> 0	<b>S</b> 0	S0	<b>S</b> 1	<b>S</b> 0
24	<b>S</b> 1	<b>S</b> 1	<b>S</b> 2	<b>S</b> 2	S2	Off	Off

## 2.2. Data processing, cleaning, and analysis

The data from the measurement campaign was treated in Python 3.8/Spyder. The data cleaning consisted of the following steps; calculating missing data, removing Not-a-Numbers (NaNs), resampling the timestamp, and interpolating. As earlier mentioned,  $O_3$  and  $NO_2$  sampled instantaneous values every 1 minute. The  $CO_2$  concentration, air temperature, relative humidity, and ventilation air flow rate) samples data based on instantaneous event changes down to 1-minute (event-based sampling) with individual timestamps. The event-based sampling occurred up to 30 times per hour for the  $NO_2$  and  $O_3$  concentrations, and as low as five times per hour. The  $CO_2$  concentration, air temperature (T), and RH were sampled approximately 8-12 times per hour. Moreover, as mentioned earlier, the ventilation air flow rate is controlled to modulate between a predefined  $V_{min}$  and  $V_{max}$ , resulting in fewer sampled data points, thus resulting in an uneven timestamp for the entire data set. The resampling was performed every 5<sup>th</sup> minute, meaning averaging each parameter every five minutes. Some minor technical difficulties occurred during the measurement campaign (e.g., unplugging the equipment and electricity breakout at the school), resulting in missing data. The high frequency of event-based sampling and resampling of the data set naturally induced some NaNs, thus linear interpolation was performed.

To find the most representative measurement of the room parameters  $CO_2$ , room temperature, and relative humidity, we performed correlation analysis using Spearman's rank correlation coefficient ( $r_s$ ) of the available measurements from the different loggers. It was decided to use the average value of the measurements for further analysis. High correlations were found for all three parameters; air temperature

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( $r_s > 0.9$ ), CO<sub>2</sub> concentration (rs=0.62 - 0.95, lowest was found for the logger placed at the window sill), and RH ( $r_s > 0.98$ ).

Violin plots were used to compare the distribution of the measurements for the different ventilation scenarios. A violin plot summarizes a data set using a boxplot to show the minimum, 1<sup>st</sup> quartile, median, 3<sup>rd</sup> quartile, and maximum and a probability density function (PDF) to show the shape of the data set. The width of the PDF indicates the frequency at which the value occurs.

#### 3. Results and discussion

In the following section, when concentration ranges are described, they refer to concentrations inside the classroom unless otherwise stated.

### 3.1. Measurement campaign results

The measured air temperature inside the classroom was relatively stable during the sampling period (weekends omitted), with daily average temperatures ranging from 22.3 °C to 23.9 °C. The highest air temperature was measured on May 30<sup>th</sup> (25.6 °C), where the outdoor temperature was 30.5 °C. Relative humidity (RH) indoors fluctuated more, with daily average values between 24.1% and 49.9%, which is lower than anticipated. During the measurement period, it was unusually sunny with relatively low daytime outdoor relative humidity (average daily RH between 35% - 83.7%). The average daily CO<sub>2</sub> concentrations varied between 460 ppm – 645 ppm, with peak concentrations between 592 ppm and 1355 ppm.

The daily average NO<sub>2</sub> concentration varied between 2.2 and 11.8  $\mu$ g/m<sup>3</sup>, with daily peak concentrations of 3.3 – 37.5  $\mu$ g/m<sup>3</sup>, which are well below the WHO recommended guidelines for indoor air of 200  $\mu$ g/m<sup>3</sup> (1-hour mean) [13]. The air quality guidelines (AQG) for short-term exposure to NO<sub>2</sub> were updated in 2021 and recommend a level of 25  $\mu$ g/m<sup>3</sup> (24-h average allowing 3 – 4 exceedance days per year)[14]. The daily average O<sub>3</sub> concentrations varied between 8.2 and 24.9 ppb (~16.5 – 49.7  $\mu$ g/m<sup>3</sup>), with peak concentrations of 16.9 – 47.1 ppb (~33.8 – 95.2  $\mu$ g/m<sup>3</sup>). The recommended short-term daily maximum AQG level for ozone is set to 100  $\mu$ g/m<sup>3</sup> (8-h average, allowing 3-4 exceedance days per year). A comparison of the daily average O<sub>3</sub> concentrations in the supply air and indoors resulted in indoor-outdoor (I/O) ratios between 0.43 – 1.02. Due to the unusually hot sunny weather during the measurement campaign, the ozone concentrations were higher than usual in Oslo during summer.



Figure 1. Correlation plot of the measured parameters, Spearman's ranked correlation.

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In the absence of indoor sources, indoor  $O_3$  concentrations are estimated to be 30-70% of outdoor levels, depending on the air exchange rate, ozone removal rate, and the reactions of ozone with other chemical compounds [10,15].  $O_3$  greatly impacts indoor air pollutants, as it can react with certain volatile organic compounds (VOCs), and the products of these reactions may be of higher concern than the VOCs themselves [9,16]. As seen in Figure 1, we found a moderate to high correlation between NO<sub>2</sub> (Spearman's ranked  $r_s$ =0.49) and  $O_3$  (Spearman's ranked  $r_s$ =0.82) with the supply airflow rate. Our findings support the notion that indoor  $O_3$  varies with the ventilation rate.

#### 3.2. Measured concentrations during occupied hours

To prevent the DCV-system from operating with maximum airflow rates during the day (unoccupied hours), we adjusted the temperature setpoint to 25 °C. Table 3 shows descriptive statistics during the hours the classroom was assumed occupied.

Parameter	Scenario	Mean	Standard deviation	Median	Minimum	Maximum
0	S0	25	6	24	12	44
0 <sub>3</sub> [ppb ]	<b>S</b> 1	20	7	19	7	38
	S2	15	6	15	6	30
NO	<b>S</b> 0	14	5	13	6	37
$INO_2$	<b>S</b> 1	12	5	11	4	30
[µg/m]	S2	9	5	10	2	26
CO	<b>S</b> 0	575	76	574	430	747
[nnm]	<b>S</b> 1	759	179	756	419	1356
[ppm]	S2	768	158	781	412	1253
Committee ainflorer note	S0	936	166	1024	429	1077
Supply alfilow rate	<b>S</b> 1	214	143	150	137	1062
[111 / 11]	<b>S</b> 2	180	87	150	138	914
$\mathbf{T}(\mathbf{n}_{0},\mathbf{n}_{0})$	SO	22.4	0.4	22.4	21.7	23.8
	<b>S</b> 1	24.2	0.5	24.1	23.1	25.6
[ C]	S2	23.3	0.6	23.4	21.1	24.3
π	S0	22.4	0.6	22.4	21.2	23.7
I (extract air)	<b>S</b> 1	24.5	0.8	24.4	22.7	26.9
[ C]	S2	23.6	0.9	23.7	20.7	25.4
RH (room) [%]	<b>S</b> 0	36	10	35	20	52
	<b>S</b> 1	40	6	41	29	52
	S2	36	8	36	19	52
DU (avtract cir)	S0	34	9	33	21	52
Kri (extract alf)	<b>S</b> 1	39	6	40	28	51
[ 70 ]	S2	35	8	36	21	53

Table 3. Descriptive statistics of the various indoor parameter during occupied hours (8:30 – 15:00).

As mentioned previously, actual detection of occupancy was performed manually and the hours the classroom was fully occupied were fewer than assumed. With the baseline  $CO_2$  setpoint of 500 ppm, the mean airflow rate during occupancy was generally closer to  $V_{max}$ , resulting in the lowest average  $CO_2$  concentrations, air temperature, and RH, but higher  $O_3$  and  $NO_2$  concentrations. In contrast, for the S1 and S2 ventilation scenario, the lower supply airflow rates due to the higher  $CO_2$  setpoint resulted in  $CO_2$  peak concentrations above 1200 ppm. These short  $CO_2$ -peak episodes (5 – 30 minutes) took place during lunch breaks or teaching sessions when the class was fully occupied. However, the lower supply air flow rate also resulted in lower  $O_3$  and  $NO_2$  concentrations. A low ventilation rate and a high air

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temperature are generally not a good combination [17], although the  $CO_2$  levels are well within the bounds where school performance would be negatively influenced [18].

## 3.3. Effect of reducing V<sub>min</sub>

As seen in Figure 2, indoor concentrations of  $NO_2$  during the nighttime with the ventilation on (S3) were comparable to when the ventilation was off. The elevated levels when the ventilation is off were most likely due to infiltration from outdoor air since there are no known emission sources of  $NO_2$  indoors. An increase in  $NO_2$  concentrations occurs first around 7 o'clock, which is during the traffic rush hours.



**Figure 2.** Average (min, max) concentration of  $O_3$  and  $NO_2$  for hours before occupancy for the three ventilation scenarios. Ventilation is usually turned on at 6:30 in the morning. Please note that the scales of the y-axis differ.

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As seen in Figure 3, the S3 scenario resulted in the lowest indoor NO<sub>2</sub> concentrations. One of the major sources of NO<sub>2</sub> is road traffic, and NO<sub>2</sub> concentrations are generally higher during winter months compared to summer. Previously, NO<sub>2</sub> was measured for five weeks during winter months inside two classrooms at the same school, and mean concentrations between  $15.4 - 45.9 \,\mu\text{g/m}^3$  were reported, with peak concentrations of approximately  $50 \,\mu\text{g/m}^3$  [19]. Ventilation is usually turned off in the afternoon and switched on 2-3 hours before occupancy. With the ventilation off, O<sub>3</sub> concentrations have a relatively rapid decay and drop to approximately zero after three hours due to the removal by indoor surfaces or reactions with VOCs. Thus, as shown in Figure 2, when the ventilation is turned on, a prominent increase in O<sub>3</sub> concentrations can be seen for both the S0 and S1 scenarios. Moreover, lowering the V<sub>min</sub> to 0.7  $1/\text{s/m}^2$  resulted in 35% lower O<sub>3</sub> concentrations on average. In contrast, if the ventilation is on during nighttime, the average indoor O<sub>3</sub> concentrations remained relatively constant at 14 ppb (~28  $\mu\text{g/m}^3$ ).



**Figure 3.** Violin plot of  $O_3$  (left) and  $NO_2$  (right) before occupancy (06:30 – 08:30) for the three ventilation scenarios. The white dot inside the box is the median.

Figure 3 shows that the S1 scenario results in the lowest median  $O_3$  concentration before occupancy and S0 the highest. The reaction of ozone with certain VOCs, i.e., terpenes, to produce secondary organic aerosols (SOA) has been well studied. Typical SOAs that are initially produced are ultrafine particles, but would with time grow into larger particles, albeit still relatively small (<700 nm) [15,20,21]. Consequently, ozone would have a great impact on the indoor level of air pollutants. The level of  $V_{min}$ and the ventilation duration would thus influence the exposure before occupancy.

#### 4. Conclusion and future outlook

In this study, different ventilation scenarios were evaluated in a classroom with demand-controlled ventilation for three weeks, focusing on exposure before occupancy. We found that reducing  $V_{min}$  to 0.7 l/s/m<sup>2</sup> would lower the ozone concentration by 35% compared to the baseline scenario with a  $V_{min}$  of 2.0 l/s/m<sup>2</sup>. Moreover, continuous ventilation during night time with a low  $V_{min}$  resulted in almost as high O<sub>3</sub> concentrations as the baseline ventilation scenario. For NO<sub>2</sub>, continuous ventilation during night time with a low  $V_{min}$  resulted in the lowest concentrations. As O<sub>3</sub> reacts easily with certain VOCs to produce secondary organic aerosols, the level of  $V_{min}$  and the ventilation duration would impact the indoor air quality upon entering the classroom. Further analysis of the VOC measurements and particle number concentrations would elucidate which ventilation strategy is most beneficial in terms of indoor air quality and energy use.

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## References

- [1] Mysen M, Berntsen S, Nafstad P and Schild P G 2005 Occupancy density and benefits of demand-controlled ventilation in Norwegian primary schools *Energy Build*. **37** 1234–40
- [2] Wachenfeldt B J, Mysen M and Schild P G 2007 Air flow rates and energy saving potential in schools with demand-controlled displacement ventilation *Energy Build*. **39** 1073–9
- [3] Merema B, Delwati M, Sourbron M and Breesch H 2018 Demand controlled ventilation (DCV) in school and office buildings: Lessons learnt from case studies *Energy Build*. **172** 349–60
- [4] Seppänen O A 2008 Ventilation Strategies for Good Indoor Air Quality and Energy Efficiency *Int. J. Vent.* **6** 297–306
- [5] Norbäck D, Nordström K and Zhao Z 2013 Carbon dioxide (CO2) demand-controlled ventilation in university computer classrooms and possible effects on headache, fatigue and perceived indoor environment: an intervention study *Int. Arch. Occup. Environ. Health* **86** 199–209
- [6] Holøs S B, Yang A, Thunshelle K and Mysen M 2019 Effect of ventilation on perceived air quality in 18 classrooms *IOP Conf. Ser. Mater. Sci. Eng.* **609** 042038
- [7] Mysen M, Holøs S, Yang A, Thunshelle K and Schild P 2019 What Should the Minimum Ventilation Rate Be in a Demand-Controlled Ventilation Strategy? *Cold Climate HVAC 2018* Springer Proceedings in Energy ed D Johansson, H Bagge and Å Wahlström (Cham: Springer International Publishing) pp 339–49
- [8] Fanger P O 1988 Introduction of the olf and the decipol units to quantify air pollution perceived by humans indoors and outdoors *Energy Build*. **12** 1–6
- [9] Salonen H, Salthammer T and Morawska L 2018 Human exposure to ozone in school and office indoor environments *Environ*. *Int.* **119** 503–14
- [10] Fadeyi M O 2015 Ozone in indoor environments: Research progress in the past 15 years Sustain. Cities Soc. 18 78–94
- [11] Standard Norge 2012 NS 3701:2012 Criteria for passive houses and low energy buildings Nonresidential buildings
- [12] Andersen K H, Holøs S B, Yang A, Thunshelle K, Fjellheim Ø and Jensen R L 2020 Impact of Typical Faults Occurring in Demand-controlled Ventilation on Energy and Indoor Environment in a Nordic Climate E3S Web Conf. 172 09006
- [13] WHO 2010 WHO guidelines for indoor air quality: selected pollutants (Copenhagen, Denmark: World Health Organization)
- [14] World Health Organization 2021 WHO global air quality guidelines: particulate matter (PM2.5 and PM10), ozone, nitrogen dioxide, sulfur dioxide and carbon monoxide (World Health Organization)
- [15] Weschler C J and Carslaw N 2018 Indoor Chemistry Environ. Sci. Technol. 52 2419-28
- [16] Wolkoff P 2020 Indoor air chemistry: Terpene reaction products and airway effects Int. J. Hyg. Environ. Health 225 113439
- [17] Wargocki P, Porras-Salazar J A and Contreras-Espinoza S 2019 The relationship between classroom temperature and children's performance in school *Build. Environ.* **157** 197–204
- [18] Wargocki P, Porras-Salazar J A, Contreras-Espinoza S and Bahnfleth W 2020 The relationships between classroom air quality and children's performance in school *Build. Environ.* **173** 106749
- [19] Yang A, Nikolaisen K F, Holøs S B, Thunshelle K, Dauge F R and Mysen M 2019 Effect of filter type in ventilation systems on NO2 concentrations in classrooms *Cold Climate HVAC 2018*. *CCC 2018* Springer Proceedings in Energy Cold Climate HVAC 2018 (Kiruna, Sweden: Springer) pp 911–21
- [20] Weschler, C.J. 2000 Ozone in Indoor Environments: Concentration and Chemistry Indoor Air 10 269–88
- [21] Fischer A, Ljungström E, Hägerhed Engman L and Langer S 2015 Ventilation strategies and indoor particulate matter in a classroom *Indoor Air* **25** 168–75